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## Original Article

# Effects of Heavy Metals on the Activity of Catalase and Glutathione-S-Transferase in Nile Tilapia Fish (Oreochromis niloticus)

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**Abstract:** This study evaluated the response of antioxidant enzymes, such as Catalase (CAT) and Glutathione-S-transferase (GST) in freshwater Nile tilapia fish *Oreochromis niloticus* (*O. niloticus*) exposed to heavy metals (HMs) including copper (Cu), lead (Pb) and cadmium (Cd). Fish were expose to various concentrations of  $Cu^{2+}$ , Pb<sup>2+</sup> (0, 0.02, 0.05, 0.2 mg/l) and Cd<sup>2+</sup> (0, 0.005, 0.01, 0.05 mg/l) for 15, 30, 45 and 60 days. The results indicated that enzyme activity was varied according to the exposure time, concentration and type of heavy metals. CAT activity increased significantly beginning at day 45 of HMs exposure. After 60 days of exposure, CAT activity was steady with Cu and Pb but inhibited by Cd. While, GST induction was earlier observed from day 15 of HMs exposure. The increase of GST activity was found with the increase of exposure time in the treatment with Cu and Cd but not with Pb. Interestingly, GST activity was inhibited by Pb at longer exposures (45 and 60 days). Among tested metals, Cu has weaker effects on the activity of CAT and GST in comparison with Pb and Cd suggesting that these enzymes were less sensitive to Cu than other tested metals.

Keywords: Heavy metal, Catalase, Glutathione-S-Transferase, Oreochromis niloticus.

#### **1. Introduction**

Copper, lead and cadmium are the most hazardous HMs that found in both aquatic and terrestrial ecosystems [1]. HMs enter, bioaccumulate and cause toxicological effects in living organisms [2]. It has been reported that HMs affect components of the cells such as lysosomes, mitochondria, nuclei and enzymes,... causing neurotoxicity, cellular function loss, cell damage and carcinogenesis [1]. These effects were used as biomarkers for metal exposure and their toxicity. Metal toxicity induces the production of free radicals which DNA damage, leads to alteration of homeostasis and stimulate lipid peroxidation [3]. Living organisms were protected from these stress by activating antioxidant defence systems [2]. GST and CAT are two important antioxidant enzymes which have been

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extensively used as biomarkers for metal exposure [4].

In waterbodies, HMs come from domestic, industrial, agricultural and other human activities [5], accumulate in aquatic organisms and affect not only their growth, development, reproduction but also on the health of human through the food chains [1, 5]. Among aquatic species, fish plays an important role in energy transfer and have higher HMs accumulation due to its high level in the food web. HMs accumulation in fish is different according to species, organs and type of metals [6, 7]. The freshwater Nile tilapia fish (O. niloticus) is an important species for commercial products in Asian countries including Vietnam. Our previous study showed that the accumulation of Cu. Cd and Pb were found in O. niloticus sampling from some lakes in Hanoi, Vietnam was site-dependent [7]. An association between the alteration of GST activity and metal accumulation was found in O. niloticus collecting in Nhue-Day river basin [8].

Therefore, this study investigated the activity of GST and CAT enzymes in *O. niloticus* exposed to Cd, Cu and Pb in various exposure periods in order to evaluate the potential effects of HMs on the antioxidant defence responses and physiological consequences of *O. niloticus*.

#### 2. Methodology

Nile tilapia fish (O. niloticus) (weight of  $7.81 \pm 1.31$  g and age, 60-70 days) were purchased from the Research Institute for Aquaculture No.1 (Bac Ninh, Vietnam). Fish were acclimated under laboratory conditions for ten days prior to the experiments. Fish were fed with commercial food twice per day at daily rate of 3-4% body weight throughout the experiment [9]. CuSO<sub>4</sub>,  $Pb(NO_3)_2$ and  $Cd(NO_3)_2$  were used as test substances which were prepared in tap water to obtain the following final dissolved concentrations: 0, 0.02, 0.05, 0.2 mg/l Cu<sup>2+</sup> or Pb<sup>2+</sup>; or 0, 0.005, 0.01, 0.05 mg/l Cd<sup>2+</sup>. These concentrations of HMs were lower than the regulation levels of National technical regulation on surface water quality (QCVN 08:2008/BTNMT). For the enzyme activity test, 40-45 acclimatized fish were distributed randomly into tanks (100L) the above mentioned which contained concentrations of HMs. At the end of the exposure period (0, 15, 30, 45, 60 days), five fish from each group were randomly taken out, dissected and their livers were collected into 2 mL Eppendorf tubes containing 500 µl Dulbecco's Phosphate Buffered Saline (DPBS) and then stored at -80 °C for enzyme activity analysis. Liver samples were defrosted on ice, homogenized and centrifuged twice at 9700 rpm for 15 min at 4 °C. Supernatants were collected for the enzymatic assay using a Sciencetific<sup>TM</sup> Thermo Biomate spectrophotometer. The CAT activity was determined following the previous method of Aebi et al. with some modifications [10]. The reaction was started by mixing 0.5 ml of UV assay substrate solution (20 mM H<sub>2</sub>O<sub>2</sub>) with 0.02 ml sample and 0.48 ml assay buffer (0.1 M K<sub>2</sub>HPO<sub>4</sub> and 0.1M KH<sub>2</sub>PO<sub>4</sub>, pH 7.0). The absorbance was measured for 30 seconds at 240 nm. The specific activity of CAT was calculated and expressed as units/min/mg protein. The GST activity was measured according to Habig et al., using 1- chloro-2,4dinitrobenzene (CDNB) as a substrate [11]. The reaction was started by mixing 0.98 mL reaction buffer (100 mM DPBS buffer (pH 6.5), 200 mM GSH and 100 mM CDNB) with 0.02 mL sample. The absorbance was measured every one minute for 8 min at 340 nm. The specific activity of GST was determined and displayed as µmoles of GSH-CDNB conjugate formed/min/mg protein. Data were processed by Excel software and statistical analyses were performed using two-way ANOVA (Bonfferroni post-tests) and presented by mean  $\pm$  SEM (n=5) using GraphPrism software.

#### 3. Results and Discussion

The liver has better oxidative stress resistance and contains higher antioxidant

enzymes than any other tissues and is usually recommended as an environmental indicator of pollution and its toxicity. Pollutant exposure stimulates the formation of oxidative stress and prompt antioxidant enzymes act as a defence mechanism in organisms. Therefore, enzyme activities in the liver are considered as sensitive biomarkers of hazardous effects from pollutants including HMs in waterbodies [4].

#### 3.1. The Effect of HMs on CAT Activity

CAT is a primary antioxidant defense component, which prevents oxidative stress damage by decomposing hydrogen peroxide into oxygen and water [4]. CAT activity (means  $\pm$  SD) in the liver of *O. niloticus* was periods. different among exposure concentrations, and types of metals (Figure 1). These activities ranged from 7.18  $\pm$  1.08 to  $93.01 \pm 14.06$ ,  $16.05 \pm 2.52$  to  $57.67 \pm 18.97$ , and  $10.32 \pm 3.02$  to  $89.07 \pm 14.35$ , in response to Cu, Pb, and Cd, respectively. The increasing trend in CAT activity according to concentration and exposure time was observed in Cu treatment. This activity increased significantly after 45 and 60 days of exposure at the concentrations of 0.05 (p<0.05) and respectively. 0.2 mg/l (p < 0.001),Pb significantly increased CAT activity from

0.02 mg/l upward after 45 (p < 0.01) and 60 days (p < 0.001) of exposure. This activity was increased continuously at 0.05 mg/l of Pb for 45 days (p < 0.05). After that, CAT activity was not change when the concentration and exposure time of Pb increased, suggesting saturation occurred in the enzyme activity with these conditions. These data showed that the CAT activity was enhanced due to the increase in Pb concentration and exposure time.

Differing from Cu and Pb, Cd showed a strange trend in CAT activity which increased on days 15 and 45 and was inhibited on days 30 and 60 of Cd exposure. However, CAT was raised insignificantly (p>0.05) and then reduced slightly at a concentration of 0.05 mg/l (p < 0.01) of Cd exposure after 15 days. The inhibition of CAT activity was found at both of the high and low concentration and short and long Cd exposure times could be explained by the involvement of Cd accumulation, toxicity and detoxification, which causes a disturbance in the body and the synthesis of enzymes [4]. The effect of the CAT activity in O. niloticus was similar to that in common carp (Cyprinus carpio) and major carp (rohu Labeo rohita) by Cd treatment [12].

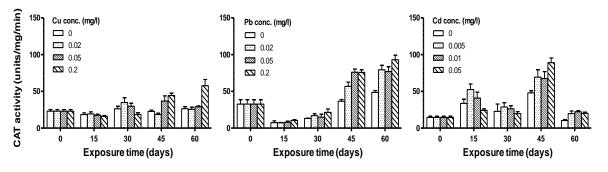


Figure 1. Liver CAT activity of *O. niloticus* exposed to HMs in different periods. Data are expressed as mean  $(n=5) \pm SEM$ .

#### 3.2. The effects of HMs on GST activity

GST is an important antioxidant enzymes which protects organisms from oxidative stress damage by catalyzing the conjugation of glutathione with metals [4]. The GST activity (means  $\pm$  SD) in the liver of *O. niloticus* also showed clear changes according to the metal concentration, exposure time and type of metals (Figure 2). These activities were 0.186  $\pm$  0.051 to 1.280  $\pm$  0.396, 0.025  $\pm$  0.010 to 0.121  $\pm$  0.020, and 0.024  $\pm$  0.008 to 0.891  $\pm$  0.286 corresponding to the Pb, Cu and Cd exposure, respectively. In general, the changes of GST activity occurred sooner than changes in CAT activity. Cu induced the GST at 0.2 mg/l after 30 days of exposure (p < 0.01). The longer the exposure time with Cu (60 days) the higher increase in GST activity, even at a low concentration of 0.02 mg/l (p < 0.001). On the 60<sup>th</sup> day of exposure, the GST activity was insignificantly increased in response to the increasing in Cu concentrations. Differing from CAT, GST activity was enhanced strongly by Pb from an earlier period (day 15) at 0.2 mg/l (p < 0.001). The increasing trend was presented

in the first 30 days at 0.05 mg/l (p<0.05). Interestingly, this trend dropped to the level as before exposure to Pb at later periods (days 45 and 60). Compared with CAT, Cd exposure enhanced (p<0.001) GST activities at the all concentrations and the exposure periods. The increase of GST activity represented the function of GST activity in oxidative stress and in protecting *O. niloticus* from damages by Cd exposure when CAT was repressed (at 0.05 mg/l). The GST activity of *O. niloticus* was similar to that of common carp (*Cyprinus carpio*) exposed to Pb and Cd [13]. GST activity was time dependent but not dose dependent in response to HMs treatment.

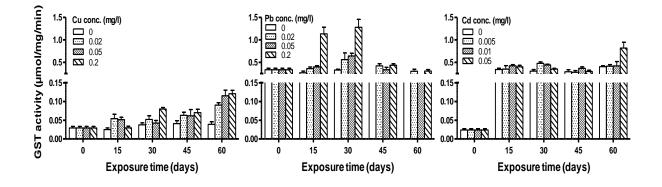


Figure 2. Liver GST activity of *O. niloticus* exposed to Pb, Cu and Cd in different periods. Data are expressed as mean  $(n=5) \pm SEM$ .

# 3.3. Correlation between Enzymatic Activity and Bioaccumulation

Our data displayed not only the activation but also the repression of enzyme activity by HMs exposure, suggesting that the activity changes might be related to the accumulation of HMs in the organism. Many scientists have been concerned about the relationship between HMs accumulation and enzyme activity in various species [8, 14, 15]. To examine whether the accumulation of Cu, Pb, and Cd affects alteration in the activity of CAT and GST, the Pearson's correlation test for the relationship between enzyme activity and HMs accumulation in muscle was performed (data not shown). The results of the relationships are shown in Table 1. The correlation analyses indicated that changes in CAT activity were related to changes in accumulation of Pb  $(r \approx 0.66, p \approx 0.002)$  and  $Cu(r \approx 0.56, p \approx 0.002)$  $p \approx 0.011$ ). Meanwhile, GST showed a positive correlation (r  $\approx$  0.81, p < 0.0001) for Cu and a weaker correlation for Cd (r  $\approx 0.55$ , p  $\approx 0.012$ ). These correlations were not tight because the HMs accumulation in muscle was tested. Therefore, the accumulation of HMs in other organs with high accumulation potential should studied. Collectively, be further the accumulation of HMs in the muscle is also an additional factor to the metal-specific enzyme activity alteration. The data showed that Cu accumulation in muscle was correlated with both CAT and GST of O. niloticus in laboratory conditions. Our findings contradicted with previous publication, which showed that Cu accumulation has no correlation with GST activity in different tissues of this species collected from the Nhue-Day River [8]. However, our results were in line with Cu and Pb but not with Cd treatment for Brown Mussels (Perna perna), which showed a high correlation (R>0.92, p<0.01) of the CAT in tissues with activity the metal Together accumulation [15]. with these findings, our data might suggests enzyme changed not only by the activity was accumulation of metal but also by other factors such as the form of metal chemicals. characteristics of organisms, and the living environment as reported in previous study [4, 12, 13]. The present data showed variability of baseline values of CAT and GST in O. niloticus which was demonstrated by the abnormality of enzyme activity without the HMs treatment in some periods. This phenomenon was also observed in common carp (Cyprinus carpio) and major carp (rohu Labeo rohita) species in our previous publications [12, 13]. These results might be due to abiotic environmental factors or the age of the fish [4].

Metals	Biological parameters			
	CAT		GST	
	R	p	R	p
Pb	0.6553	0.0017	0.2303	0.3286
Cu	0.5566	0.0108	0.8093	< 0.0001
Cd	0.2821	0.2282	0.5520	0.0116

Table 1. Pearson correlation coefficients between HMs accumulation and enzyme activity

3.4. Comparison of the Sensitivity of Enzymes with HMs

To better understand the response of enzymes in O. niloticus to HMs exposure, the activity of CAT and GST were compared at the same concentration (0.05 mg/l) of different metals in several exposure periods. Among the test compounds, Cu, which is an essential metal in organisms for enzymes to function normally, showed the lowest effect on both CAT and GST in comparison with other metals (Figure 3). During a short exposure time (day 15), there was an insignificant change in the activity of CAT and GST at all concentrations. Our results agreed with the results of a previous report that showed CAT activity in the liver of O. niloticus did not alter significantly at all tested concentrations of Cu [16].

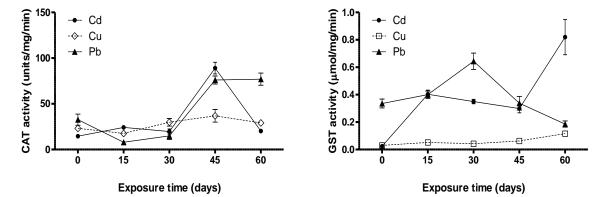


Figure 3. Comparison of enzymatic activities in livers of *O. niloticus* exposed with the same concentration (0.05 mg/l) of Pb, Cu and Cd in different periods. Data are expressed as mean  $(n=5) \pm SEM$ .

#### 4. Conclusion

Our data revealed that the enzyme activities of O. niloticus exposed to Cu, Pb, and Cd varied depending on the concentrations, exposure times and types of metals. Cu showed a weaker effect on the activity of CAT and GST in comparison to Pb and Cd, suggesting that these enzymes were less sensitive to Cu than other tested metals. Our findings also suggest that CAT and GST are sensitive biomarkers for metal biomonitoring the in aquatic environment. The elimination time of HMs, which is also an important factor for enzyme changes, should be a concern in future studies.

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